Report on Monitoring of Forest Changes in Pin Supu Sustainable Forest Management Project Area

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SABAH FORESTRY DEPARTMENT

1. INTRODUCTION

Pin Supu Forest Reserve (PSFR) is a Class VI Protection Forest with a total area of 4,696 ha. The Kinabatangan District Forestry Office administers the reserve. PSFR consists of three units of land separated by highways and rivers at considerable distances. In the mosaic landscape of the Kinabatangan, Blocks A and B border onto other protected areas, such as Lot 8 in the west and Lot 7 of the Wildlife Sanctuary. However, Block C is entirely isolated and surrounded by oil palm estates (Figure 1).

On a landscape level, the PSFR is part of the Kinabatangan floodplain ecosystems (Nilus *et al.*, 2015). The entire area is below 200 m in elevation and consists of secondaryand advanced-growth of mixed dipterocarp, seasonal freshwater and freshwater swamp forests. Fauna assessment has shown that diverse fauna can be found within the area. The area provides habitats for the fauna and acts as a transient wildlife migratory path between the different forest reserves it borders. On the cultural aspect, the community cooperative initiative (KOPEL) was given the right to harvest bird nests at Supu Caves. Hence, the continuous food source for swiftlet (insects) and the presence of forest surrounding the limestone caves is significant.

Through the Forest Research Centre, the Forestry Department has established 10 (0.13 ha) permanent sample plots and conducted three censuses: September 2014, November 2017 and August 2023 (Figure 2; Table 1). This report evaluates forest change after nine years of monitoring period by the department, and the information will provide forest ecosystem background in the various forest types found in PSFR.

2. OBJECTIVES

The objectives of the monitoring activities are to investigate changes in trees ≥ 10 cm DBH in the following details:

- i. Forest cover changes between 2008, 2017 and 2023
- ii. Plot similarities in species composition
- iii. Mortality and recruitment rates
- iv. Growth
- v. Above-ground biomass

Figure 1. The location map of the Pin-Supu Forest Reserve in Sabah, Malaysia.

Figure 2. The location of the ten permanent sample plots in Pin-Supu Forest Reserve, Sabah, Malaysia.

3. METHODOLOGY

3.1 Forest Cover Changes Pin-Supu Forest Reserve

The assessment of forest cover dynamics within the Pin-Supu Forest Reserve was quantified using remote sensing. Six satellite image scenes were selected based on acquisition years, mainly 2008, 2016, 2020 and 2023 (Table 2).

Satellite	Year	Path/Row	Acquisition Date	Cloud coverage (5%)
Landsat 5 TM	2008	117/56	20 May 2008	Less than 30%
Landsat 8 OLI	2015	117/56	12 August 2015	Less than 30%
	2016	117/56	13 July 2016	Less than 30%
	2020	117/56	26 September 2020	Less than 30%
	2023	117/56	17 July 2023	Less than 30%

Table 2. Satellite image acquired for the forest cover change detection.

Data Acquisition and Pre-Processing

All satellite images were acquired and downloaded from the USGS (glovis.usgs.gov). The images were radiometrically pre-processed in QGIS (Quantum GIS) to convert the DN values into reflectance values.

Each scene was further processed in ArcGIS Pro for classification. Six land covers were identified: water bodies, dryland forest, swamp vegetation, shrubland, oil palm and bare land. A composite image comprising band 5, band 4 and band 3 for Landsat 5 was used in the classification. A composite of Band 6, Band 5 and Band 2 was used for Landsat 8 OLI. At least 15 training samples were collected for each land cover. Training sample selection was based on ancillary data such as elevation data, natural vegetation maps, soil maps and observation from high-resolution imagery on Google Earth. The image was classified using pixel-based supervised classification and support vector machine (SVM) as the classifier.

The classification for the forested area was divided into three sub-classes based on the spectral variation of the tree crowns that can be observed, such as light green, medium and dark green-coloured tree crowns. These sub-classes were then merged into one class and classified as dryland forests. Based on the soil map, natural vegetation map and soil wetness index, forested areas located below 30 m (a.s.l) are classified as swamp vegetation. The swamp vegetation was also divided into three sub-classes based on the spectral variation of vegetation: bright, medium and brown. These sub-classes were subsequently merged into a single "swamp vegetation" class to represent a broader characterisation of vegetation in swampy areas. The classified image was then post-processed using a 3 x 3 majority filtering to reduce the 'salt and paper' effect. Any misclassified pixels of swamps in areas where the elevation is more than 30 m (a.s.l) were reclassified as dryland forests, given the normality of peat swamps to be on flat sites. The classified raster was then converted to vector data. Dryland and swamp vegetation were delineated by identifying wet areas using MNDWI (Modified Normalized Difference Water Index).

3.2 Re-measurement of recorded trees

All previously labelled trees ≥ 10 cm diameter at breast height (DBH) were re-measured, and their species identity was re-confirmed. Newly recruited trees ≥ 10 cm diameter at breast height were labelled, measured and identified to species level.

3.3 Data manipulation and analysis

The data set within each sample plot was checked for anomalies such as abrupt changes in an individual's size or irreconcilable changes in species identities (different families or genera). All anomalies were rectified to avoid excluding data from the sample.

All tree data were subject to Bray-Curtis Ordination using the R statistical package to investigate plot differences based on dissimilarity in species composition and their abundance.

Mean annual mortality rates (*m*) were estimated using the equation provided by Sheil *et al*. (1995):

 $m = 1 - (N_1 / N_0)^{1/t}$

*N*₀ is the number of trees at the beginning of a census interval, and *N*₁ is the number of trees surviving at the end of the census interval *t* (years).

Mean annual recruitment rates (*r*) were calculated using the equation provided in Sheil (1996):

 $r = 1 - (1 - n_r / N_t)^{1/t}$

n^r is the number of recruits, and *N^t* is the number of trees at the end of the census interval *t* (years).

Turnover rates were estimated as the mean of mortality and recruitment rates.

Annual diameter increment (AGR) and relative growth rate of diameter (RGR) were calculated using the following equations:

 $AGR = (x_t - x_0) / t$

 $RGR = (x_t - x_0 / x_0) / t \times 100 \%$

*x*⁰ and *x^t* are DBH at the beginning and end of census interval *t* (years).

Potentially erroneous tree growth data were identified using the criteria adopted by Condit *et al.* (1993b). Trees that shrank by more than 5% of their initial diameter per year or exceeded a mean annual diameter increment of 75 mm per year were discarded from the analysis. These minimum and maximum thresholds for growth rates have successfully avoided growth anomalies and provided estimates close to the median of each group of growth data in other studies (Condit *et al*. 1993a). Fortunately, no trees were omitted from the growth analysis.

The aboveground carbon estimation of individual trees can be estimated from the measured diameter using the aboveground biomass allometric regression equation by Chave *et. al* (2014) that is suited and widely used by the Sabah Forestry Department:

AGB= [exp (-1.803 – 0.976E + 0.976 $ln(p)$ + 2.673 $ln(DBH_i)$ – 0.0299 $[ln(DBH_i)]²]$

E= climate variable (http://chave.upstlse.fr/pantropicalallometry.htm)

The individual tree biomass value that derives from the equation will be summed to produce the biomass of the sample plot, which is then multiplied by a standard value of carbon concentration to produce an estimate of carbon stock. An assumption value of 47% of the dry biomass is carbon.

AGB Carbon = $0.47 \times$ AGB (kg/ha)

4. RESULTS

4.1 Forest Cover Changes

The dynamic changes in forest cover within the Pin-Supu Forest Reserve were evaluated through image processes of six satellite image scenes, specifically in 2008, 2016, 2020, and 2023. The area for each land cover was calculated and depicted in Table 3. However, cloud cover was inevitable in the 2008 satellite image. Appendix I details the land cover distribution for four observed years (Figures 3–6). Over the past 15 years, forest expansion has shown an apparent upward trend in dryland (30 ha) and swampland (267 ha), as listed in Table 3. Correspondingly, there is a notable decline in the extent of marshland and shrub cover within dry areas. Water bodies exhibit an upward trend in land cover, while bare land is on a decreasing trajectory. However, the oil palm cover displays irregularities.

Table 3. The area and percentage (in parentheses) of vegetation and other land cover of Pin Supu Forest Reserve in 2008, 2016, 2020 and 2023.

4.2 Plot similarities in relation to species composition

The dendrogram reveals two distinct groupings based on the water table and soil drainage characteristics of the site, namely, inland and swamp ecosystems (Figure 2; Table 4). In regions with a moderately high-water table, the floristic composition of disturbed lowland seasonal freshwater swamp forests is evident in PSP 3, 7, and 8. In contrast, the highest water table, characteristic of freshwater swamp forests, is represented in PSP 4. PSP 1, 2, 6, and 9 are established in drier sites, featuring various regenerative lowland mixed dipterocarp forests. However, according to the soil association map, PSP 10 is situated in an area marked by a high-water table or formerly a swamp forest, currently dominated by Talisai paya trees.

Plot dissimilarity by Species

Figure 2 A dendrogram of all 10 permanent sample plots (PSPs) are clustered according to water-table condition in the Pin Supu Sustainable Forest Management project area, Sabah.

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	PSP ₁	PSP ₂	PSP ₃	PSP ₄	PSP ₅	PSP ₆	PSP ₇	PSP ₈		PSP ₉	PSP 10
PSP ₁	0	0.672	0.968	0.990	0.984	0.940	0.985	0.957		0.917	0.973
PSP ₂	0.672	0	0.968	0.981	0.968	0.940	0.955	0.957		0.917	0.973
PSP ₃	0.968	0.968	0	0.898	0.897	0.904	0.577	0.527		0.926	0.927
PSP ₄	0.990	0.981	0.898	0	0.939	0.961	0.892	0.896		0.963	0.954
PSP ₅	0.984	0.968	0.897	0.939	0	0.920	0.870	0.878		0.911	0.912
PSP ₆	0.940	0.940	0.904	0.961	0.920	$\mathbf 0$	0.909	0.857		0.833	0.863
PSP ₇	0.985	0.955	0.577	0.892	0.870	0.909	0	0.551		0.887	0.889
PSP ₈	0.957	0.957	0.527	0.896	0.878	0.857	0.551	0		0.880	0.868
PSP ₉	0.917	0.917	0.926	0.963	0.911	0.833	0.887	0.880		0	0.885
PSP 10	0.973	0.973	0.927	0.954	0.912	0.863	0.889	0.868		0.885	0

Table 4. The dissimilarity index values among 10 plots established in the Pin Supu Sustainable Forest Management project area, Sabah.

4.1 Comparison of population dynamics for trees ≥10 cm DBH among plots

4.1.1 Mortality, recruitment and turnover

A comparative analysis of tree mortality and recruitment across all sample plots spanning the two monitoring periods, 2015–2018 and 2018–2023, in the Pin Supu SFM project area is detailed in Table 5. Throughout the eight-year monitoring period, an average mortality rate of 3% (equivalent to 144 trees) and a recruitment rate of 2% (70 trees) were documented, resulting in twice as many recorded deaths as recruitments. Notably, Plot 4 and Plot 10 exhibited the highest number of tree fatalities, with mortality rates ranging from 2.03% to 7.76% per year and 4.38% to 6.55% per year, respectively.

Table 5. Summary of mortality, recruitment and turnover of trees ≥ 10 cm DBH in all 10 permanent sample plots (PSPs) between 2015–2018 and 2018–2023 in Pin Supu Sustainable Forest Management project area, Sabah, Malaysia.

4.1.2 Growth

Table 6 illustrates a comparison of tree growth across all sample plots. On average, trees with a diameter at breast height (DBH) \geq 10 cm in Plot 2, 5, and 10 have consistently shown positive growth, maintaining elevated Average Growth Rates (AGR) and Relative Growth Rates (RGR) in both monitoring periods. In contrast, trees in Plot 7 exhibited the lowest growth.

Table 6 Growth of trees ≥ 10 cm DBH in all 10 permanent sample plots (PSPs) between 2015–2018 and 2018– 2023 in Pin Supu Sustainable Forest Management project area, Sabah, Malaysia. (Annual Growth Rate, AGR; Relative Growth Rate, RGR)

4.2 Comparison of population dynamics for trees ≥10 cm DBH among forest ecosystems

4.2.1 Mortality, recruitment and turnover

A comparison of tree dynamics in three distinct forest ecosystems between the 2015 – 2018 and 2018 – 2023 periods is presented in Table 7 and 8. The Talisai Paya swamp forest demonstrated the highest turnover rate for both census periods (Table 6). The riparian forest, SFWSF, and FWSF exhibit an increasing trend in turnover rate, while the Talisai Paya swamp shows a decreasing trend throughout the monitoring period. The MDF indicates no changes in turnover rate during the census. The Talisai Paya swamp exhibits the highest mortality rates in the first monitoring period, while FWSF shows the highest mortality rates in the second monitoring period. The Talisai Paya swamp demonstrates the highest recruitment rates in both monitoring periods (Table 7). Throughout the observed ecosystems, approximately 50% and 70% of all tree deaths were represented by the lowest diameter size class (10.0 to 19.9 cm) in the first and second monitoring periods, respectively (Table 8).

Table 7 Summary of mortality, recruitment and turnover of trees ≥ 10 cm DBH observed within mixed dipterocarp forest (MDF), riparian forest, Talisai Paya swamp forest, seasonal freshwater swamp forest (SFWSF)

and freshwater swamp forest (FWSF) between 2014–2023 in Pin Supu Sustainable Forest Management project area, Sabah, Malaysia.

Table 8 Tree dead of trees ≥ 10 cm DBH based diameter classes within mixed dipterocarp forest (MDF), riparian forest, Talisai Paya swamp forest, seasonal freshwater swamp forest (SFWSF) and freshwater swamp forest (FWSF) between 2015–2023 in Pin Supu Sustainable Forest Management project area, Sabah, Malaysia.

4.2.2 Growth

A comparison of tree growth on three distinct forest ecosystems in two monitoring periods is presented in Table 9. The Talisai Paya swamp and riparian forests demonstrated high annual growth and relative growth rates in both monitoring periods. The SFWSF and FWSF exhibited

an increasing trend of AGR. However, only the former demonstrated an increasing trend of RGR. Only MDF showed consistent AGR and RGR in both monitoring periods.

Table 9 Growth of trees ≥ 10 cm DBH within mixed dipterocarp forest (MDF), riparian forest, Talisai Paya swamp forest, seasonal freshwater swamp forest (SFWSF) and freshwater swamp forest (FWSF) between 2015–2023 in Pin Supu Sustainable Forest Management project area, Sabah, Malaysia.

Forest	Mean of AGR	Mean of AGR	Mean of RGR	Mean of RGR	
Condition	2015-2018 (cm/yr)	2018-2023 (cm/yr)	2015-2018 (%)	2018-2023 (%)	
MDF	0.27	0.26	1.30	1.17	
	± 0.05	± 0.02	± 0.12	± 0.11	
Riparian	0.50	0.47	2.18	± 0.25	
	± 0.15	± 0.08	± 0.52	1.66	
Talisai Paya	0.49	0.51	2.80	2.55	
swamp	± 0.09	± 0.08	± 0.42	± 0.48	
SFWSF	0.07	0.20	± 0.12	0.92	
	± 0.03	± 0.02	0.47	± 0.09	
FWSF	0.19	0.36	± 0.28	± 0.17	
	± 0.05	± 0.05	1.04	1.38	
Grand	0.23	0.30	± 0.10	1.29	
Total	± 0.03	± 0.02	1.20	± 0.07	

4.2.3 Species compositional changes

The MDF has the highest number of species recorded in PSFR, while FWSF and Talisai Paya swamp forests have the lowest (Table 10; Appendix II). Throughout the monitoring period, the reduction in the total number of species is negligible. MDF depicts the highest number of species recorded as dead and recruited in both monitoring periods (Table 11). The composition of dead and recruited trees varies among successional groups, including mixed climax and pioneer species, as well as mixed structural canopy layers such as the main canopy, middle storey, and understorey species.

Table 10 Number of tree species with ≥ 10 cm DBH within mixed dipterocarp forest (MDF), riparian forest, Talisai Paya swamp forest, seasonal freshwater swamp forest (SFWSF) and freshwater swamp forest (FWSF) between 2015–2023 in Pin Supu Sustainable Forest Management project area, Sabah, Malaysia.

Table 11 List of species recruited and tree death \geq 10 cm DBH within mixed dipterocarp forest (MDF), riparian forest, Talisai Paya swamp forest, seasonal freshwater swamp forest (SFWSF) and freshwater swamp forest (FWSF) between 2015–2023 in Pin Supu Sustainable Forest Management project area, Sabah, Malaysia.

4.2.4 Above-ground Carbon changes

In the 2023 census, the overall average above-ground biomass (AGB) values in all plots indicate a 9% increase from the initial values in 2015 (Table 12). The recorded overall averages of AGB in all plots are approximately 349 C t/ha in 2015 and 378 C t/ha in 2023.

Two plots, PSP 3 (SFWSF) and PSP 10 (Talisai Paya forest), demonstrate an overall increase in forest structure values and also show 9% and 29% increases in initial above-ground carbon (AGC), respectively (Table 11).

The riparian forest represented by PSP 5 shows the highest increase in AGC from the initial values (36%), despite a 10% decrease in initial tree density (Table 11). Similarly, Plot 4 (FWSF), Plot 6 (MDF), and Plot 8 (SFWSF) indicate reductions in tree densities but show increasing AGC values of 23%, 13%, and 6% from the initial AGC, respectively. Only PSP 7 (SFWSF) depicts a decrease in structural values, but AGC remains consistent.

In the MDF, PSP 1 and PSP 2 demonstrate a reduction in forest structural values and a loss of 3% and 8% of their initially estimated above-ground carbon, respectively (Table 11). However, other MDF plots, specifically PSP 9, exhibit a 3% increase in initial above-ground carbon, despite a reduction in forest structural values.

Table 12 Summary of forest structure and estimated above-ground Carbon (AGC) per hectare of lived standing trees with \geq 10 cm DBH within mixed dipterocarp forest (MDF), riparian forest, Talisai Paya swamp forest, seasonal freshwater swamp forest (SFWSF) and freshwater swamp forest (FWSF) between 2015–2023 in Pin Supu Sustainable Forest Management project area, Sabah, Malaysia.

Forest condition	Plot	Forest structure details	2015	2018	2023
Advance-growth of mixed	$\mathbf{1}$	Density/ha (Stem/ha)	436.5	428.6	436.5
dipterocarp forest		BA/ha (m ² /ha)	30.2	28.9	28.7
		Volume (m ³ /ha)	370.9	355.4	353.6
		AGC (C t/ha)	389.9	377.9	378.4
Advance-growth of mixed	$\overline{2}$	Density/ha (Stem/ha)	420.6	412.7	412.7
dipterocarp forest		BA/ha (m ² /ha)	31.0	35.2	27.5
		Volume (m^3/ha)	382.1	427.1	335.9
		AGC (C t/ha)	392.3	461.1	359.8
Secondary-growth of Seasonal	3	Density/ha (Stem/ha)	428.6	428.6	428.6
Freshwater Swamp Forest		BA/ha (m ² /ha)	23.1	23.3	25.1
		Volume (m ³ /ha)	293.2	296.8	317.6
		AGC (C t/ha)	296.9	300.1	323.8
Secondary-growth of Freshwater	4	Density/ha (Stem/ha)	1063.5	1007.9	650.8
Swamp Forest		BA/ha (m ² /ha)	39.7	40.7	43.0
		Volume (m ³ /ha)	521.1	531.8	545.9
		AGC (C t/ha)	417.1	434.8	515.5
Secondary-growth of previously	5	Density/ha (Stem/ha)	381.0	349.2	341.3
mixed dipterocarp forest (Riparian Forest)		BA/ha (m ² /ha)	23.4	27.3	28.9
		Volume (m^3/ha)	295.3	338.6	356.6
		AGC (C t/ha)	224.0	279.6	305.3
Advance-growth of mixed	6	Density/ha (Stem/ha)	460.3	460.3	452.4
dipterocarp forest & limestone vegetation		BA/ha (m ² /ha)	32.6	34.1	35.9
		Volume (m ³ /ha)	400.4	417.5	437.9
		AGC (C t/ha)	437.3	462.2	494.4
Secondary-growth of Seasonal	$\overline{7}$	Density/ha (Stem/ha)	444.4	412.7	396.8
Freshwater Swamp Forest		BA/ha (m ² /ha)	24.2	23.3	24.1
		Volume (m^3/ha)	304.5	293.2	301.0
		AGC (C t/ha)	281.6	273.1	283.9
Secondary-growth of Seasonal	8	Density/ha (Stem/ha)	484.1	468.3	444.4
Freshwater Swamp Forest		BA/ha (m ² /ha)	33.2	32.2	34.2
		Volume (m ³ /ha)	413.2	400.2	422.3
		AGC (C t/ha)	474.8	459.8	501.4
Advance-growth of mixed	9	Density/ha (Stem/ha)	468.3	444.4	412.7
dipterocarp forest		BA/ha (m ² /ha)	33.8	32.3	32.6
		Volume (m ³ /ha)	416.1	395.6	394.9
		AGC (C t/ha)	390.4	391.1	401.6
Talisai Paya Swamp Forest	10	Density/ha (Stem/ha)	404.8	388.9	412.7
(formerly Seasonal Freshwater Swamp Forest)		BA/ha (m ² /ha)	20.5	20.4	23.8
		Volume (m ³ /ha)	262.0	259.2	298.2
		AGC (C t/ha)	181.1	189.3	234.2

5. DISCUSSION

5.1 Limitation of findings

In this report, data on above-ground carbon, tree growth, mortality, and recruitment rates have been collected from three censuses conducted at 8-year intervals. The monitoring results for various forest types in both the floodplain and dryland are highly dynamic, and caution should be exercised in interpreting the findings. Further long-term assessments are required to gain a comprehensive understanding of the latest monitoring results.

5.2 Changes in forest cover

Floodplain forests are dynamic ecosystems found along riverbanks that experience periodic flooding. The expansion and contraction of their extend are shaped by the interplay of water flow, sediment deposition, and vegetation dynamics. Historically, extreme natural events like floods and droughts exert selective pressure on populations, and variations in water flow patterns can influence the relative success of different species and regulate ecosystem process rates (Resh et al., 1988; Hart & Finelli, 1999).

The expansion of the forest from 687 ha to 954 ha in 15 years on wetlands is noteworthy (Table 3). The swamp trees, known for their rapid growth and turnover, likely accelerated the succession process, as indicated in the earlier findings. Additionally, some degraded swampland has been treated and rehabilitated, aiding in the regeneration process.

Approximately 96% of the dryland identified in 2008 was categorised as forest cover, while the remaining portion had shrub vegetation (Table 3). The slow progression of the dryland forest may be attributed to unfavourable soil productivity, such as skid trails, stumping points, or former campsites on the site. Additionally, the forest edge effect may have limited any succession processes.

It is envisaged that as forests grow and expand, contribution towards biodiversity conservation, and providing habitats for numerous plant and animal species could be significant. Moreover, the expanding forests could enhance its supporting services in primary production, nutrient cycling and water cycling. These high-growth and dynamic forests could act as vital carbon sinks, helping mitigate climate change by absorbing and storing carbon dioxide (MEA 2005). Moreover, these forests also play a crucial role in maintaining water quality, regulating local climates, and preventing soil erosion. Forest expansion promotes ecosystem resilience and is of paramount importance for ecological balance and human well-being.

5.3 Changes in forest dynamics and growth over time

Regeneration of residual stand forest in Pin Supu after human-induced disturbance, such as timber extraction and forest fire influenced by the level of degradation of the site, availability of regenerative seedlings or saplings, and undergrowth competition such as herbaceous climbers and sedges. The loss of large trees that were extracted during logging activities in advance-growth forests in the past creates canopy gaps that stimulate the growth of many neighbouring species of various sizes, e.g. understorey seedlings, saplings and pole-size trees (Phillips *et al.* 1994). Thus, the findings of low recruitment over mortality rate and yet complimented with positive tree growth and incremental trend of above-ground biomass may indicate that these natural forests are recuperating, hence demonstrating that the forests are on a successional trajectory towards diverse composition and structural forests.

5.4 Variation in forest dynamics growth among forest communities

The findings have delivered some evidence of differences in forest communities in their assemblages, growth and dynamics of trees \geq 10.0 cm DBH are significantly related to the edaphic conditions that are likely to be associated with the gradient of resource availability that distinguished forest communities in terms of their floristics, structure and diversity. Further monitoring is required to rationalise these observations since data censuses are based on an eight-year period only.

6. MANAGEMENT RECOMMENDATION

6.1 Protection of forest

Continue utilising stand-based mapping of vegetation types monitoring purposes in the management of this conservation area. This management tool may be able to examine spatiotemporal processes of changes in forest quality and conditions.

6.2 Maintenance of PSPs

Permanent plots require ongoing maintenance and when left unattended for long periods of time, they become increasingly difficult to relocate, re-establish, and undertake accurate remeasurements. The maintenance of permanent plots consists of determining the presence of the centre post and tree, including looking out for severe damage to the plots and investigating its cause.

6.3 Establish additional plots

Additional establishment PSPs are required in other parts of the reserve for successional comparison with the existing PSPs that were established in advance-growth forests.

7. SYNTHESIS

Two ecosystems, dryland and wetland, previously disturbed in various regenerative and successional stages, are under continuous monitoring. Over the past 15 years, there has been a noticeable upward trend in forest expansion in dryland areas (30 ha) and swampland (267 ha), accompanied by a significant decline in marshland and shrub cover within dry regions. The forests are in the process of recovering from previous disturbances, a trend evident over the eight-year monitoring period from 2014 to 2023. Positive indicators include overall tree growth, the recruitment of diverse species, and a favourable change in above-ground biomass or carbon. A long-term monitoring program for forest health is crucial to understanding the significant ecosystem services that contribute to ecological balance and human well-being.

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Appendix I Land cover map of Pin Supu Forest Reserve in 2008, 2016, 2020 and 2023.

Figure 3. Land cover map of Pin Supu Forest Reserve in 2008.

Figure 4. Land cover map of Pin Supu Forest Reserve in 2016.

Figure 5. Land cover map of Pin Supu Forest Reserve in 2020.

Figure 6. Land cover map of Pin Supu Forest Reserve in 2023.

APPENDIX II List of species recorded in various forest ecosystems

